

Response of Soil Properties and Microbial Communities to the Application of Olive Mill Wastewaters

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ABSTRACT

A laboratory trial was carried out to evaluate the short-term influence of different concentrations of olive mill wastewaters (OMW) on soil physico-chemical properties and microbial community. After 30 days' incubation, no significant changes occurred in soil pH values and humified organic carbon; soil organic carbon and total water-soluble phenols (WSP) significantly increased. Addition of OMW to the soil caused modification in microbial counts and microbial community structure. Bacteria decreased significantly up to 86% for 100% OMW treatment. No significant differences were found for fungal community in the soil treated up to 50% OMW while a decrease in 50% was found when 100% OMW was applied. The fungal: bacteria ratio increased significantly. The actinomycetes CFU (colony forming units) decreased gradually up to a maximum of 38% when OMW dose increased. Our results evidenced that the storage of OMW in open-air lagoons for a 4-month period before its soil application and the appropriate dilution with irrigation water without any further treatment can be an inexpensive technology to be adopted by small-sized olive mills.

Keywords: microbial counts, organic carbon, physico-chemical properties, phenols, soil fertility, soil quality

Abbreviations: BOD, biochemical oxygen demand; C, carbon; CFU, colony forming units; COD, chemical oxygen demand; EC, electrical conductivity; FA, fulvic acid; HA, humic acid; HR, humification ratio; PVP, polyvinylpyrrolidone; TAE, tannic acid equivalents; TOC, total organic carbon; TPP, total polyphenols; WHC, water-holding capacity; WSP, water-soluble phenols

INTRODUCTION

In Mediterranean countries, over a limited time period (usually from October to December), the olive oil industry produces large amounts of liquid and solid wastes, whose characteristics, particularly moisture and oil contents, depends on a broad variety of factors such as type and maturity of olives, harvesting time, region of origin, climatic conditions, cultivation practices and the extraction process which is employed (Azbar *et al.* 2004; Neto Andre *et al.* 2005; Roig *et al.* 2005).

Olive mill wastewaters (OMW) are formed by the tissue water from the olive fruits, the water added during the oil extraction process, soft tissues from olive pulp and oil, in the form of a very stable emulsion (Lanciotti *et al.* 2005). The removal of OMW is a crucial problem for the whole community in general and for the producers and millers in particular, because of their polluting effects on soil and water (Cegarra *et al.* 1996; Paredes *et al.* 1999; Filippi *et al.* 2002). This is due to its high organic load (Paredes *et al.* 1999), phytotoxic properties, (Casa *et al.* 2003; Kistner *et al.* 2004; Isidori *et al.* 2005), antibacterial (Pèrez *et al.* 1992) and antimicrobial (Fiorentino *et al.* 2003) activities, and relatively low biodegradability (Komilis *et al.* 2005). In order to find convenient solutions for the using of this by-product without harmful environmental effects, many studies were conducted in several olive-oil producing countries and different disposal methods based on evaporation ponds, thermal concentration, physico-chemical and biological treatments as well as direct application to agricultural soils as organic fertilizers have been proposed (Rozzi and Malpei 1996).

OMW contain substantial amounts of plant nutrients and organic carbon and may also represent a low cost source of water (Cegarra *et al.* 1996). In countries on the south bank of the Mediterranean with severe water deficient

environments and with soils usually characterized by a scarcity of organic matter, the use of such wastes for soil fertigation could be even more beneficial, although toxic effects on microbial growth and plant seed germination have been observed (Capasso *et al.* 1992). Many study explored the impact of OMW on various soils, showed a temporary decrease in pH, plant available magnesium and hydraulic conductivity, followed by an increase in soil salinity and bulk density (Saviozzi *et al.* 1991; Pèrez *et al.* 1992; López *et al.* 1996; Colucci *et al.* 2002; Saadi *et al.* 2006).

In contrast, numerous studies have demonstrated that land spreading of OMW increases soil fertility (Martens *et al.* 1992; Perucci 1992; Albich *et al.* 2000), improving soil porosity (Pagliai 1996), stabilizing soil conglomerates (Flouri *et al.* 1988; Saviozzi *et al.* 1991; Tomati and Galli 1992; Tomati *et al.* 1996; Colucci *et al.* 2002; Ferri *et al.* 2002), increasing soil organic matter, available P and exchangeable K contents (Brunetti *et al.* 1995; Ferri *et al.* 2001). In addition, Potenz *et al.* (1980), Marsilio *et al.* (1991) and Senette (1991) indicated that OMW had low potentially toxic heavy metal concentrations, so that their application to soil did not lead to environmental pollution risk.

Furthermore, even if many authors investigated the short-term effects of spreading OMW on soil microbial biomass, only limited information is available on the effects of OMW application on soil microbial community structure, which in turn may influence the viability of agriculture soils.

Paredes *et al.* (1986) found decrease in the number of *Bacillus* sp., Tardioli *et al.* (1997) evidenced changes in soil fungal compositions: these findings were confirmed by Mechri *et al.* (2007) that found an alteration in the microbial community structure after agronomic application of OMW.

Because of the existing contradictory results, not always useful to draw practical conclusion, and the limited infor-

Table 1 Chemical characteristics of olive mill wastewater fresh collected (unstored) and 4-months stored in cement tank in open air (stored).

Parameters	Un-stored OMW	Stored OMW
pH	4.5 ± 0.2	5.5 ± 0.1
BOD ₅ g L ⁻¹	8.0 ± 0.5	4.1 ± 0.8
COD g L ⁻¹	18.9 ± 0.6	10.5 ± 1.3
EC dS/m	5.1 ± 0.4	4.59 ± 0.03
Organic matter g L ⁻¹	11.5 ± 0.7	9.46 ± 1.0
Ash g L ⁻¹	5.1 ± 0.7	3.7 ± 0.5
Dry matter g L ⁻¹	11.1 ± 0.1	9.80 ± 0.8
WSP mg TAE L ⁻¹	25.5 ± 0.1	5.47 ± 0.5
TPP mg TAE L ⁻¹	31.6 ± 0.7	11.3 ± 0.7

Each value represents mean of five independent observations ± S.E.

mation on the effects of OMW application on soil microbial community structure, the present work was aimed at assessing, under laboratory conditions, the direct application of raw OMW to soil. To this end, the short-term influence of different OMW concentrations on several chemical soil properties and microbial community structure, reflecting soil ecosystem quality and functioning was studied.

MATERIALS AND METHODS

Olive mill wastewater

The OMW samples were collected from an olive oil mill located in the city of Lametia Terme (Italy), which employs a three-phase decanter centrifuge process for oil separation, and stored in cement tanks of 200 m³ in October 2006 and then taken in February 2007. In this period the meteoric contribution was 0.32 m³. The characteristics of OMW are given in **Table 1**. The OMW were refrigerated at 4°C until used.

Sampling and physico-chemical analyses of soil

Composite soil samples (0-20 cm) were taken from the Agrarian Firm of the Agriculture Faculty of Reggio Calabria. According to USDA classification, the soil can be classified as sandy loam soil (clay 16%, sand 82%, silt 2%). Soil samples were stored at 4°C until the analysis.

Soil analyses were performed on air dried and sieved (<2 mm) soil samples. Particle-size analysis was carried out by the hydrometer method using sodium hexametaphosphate (SHMP) as a dispersant (Bouyoucos 1962); pH was measured in distilled water (solid: liquid ratio 1: 2.5) with a glass electrode; organic C was estimated by the Walkley-Black procedure (Nelson and Sommers 1982) and was converted to organic matter by multiplying the percentage of C by 1.72. The total CaCO₃ was carried out using a De Astis Calcimeter. Humic substances were extracted with 0.1 N NaOH and 0.1 N Na₂P₂O₇ solution (solid: liquid ratio 1: 10); the suspension was shaken for 16 h at room temperature and centrifuged at 5,000 rpm for 30 min; the extract was dialysed by Wisking tubes against distilled water to pH 6.0. Subsequently, the solution was filtered through a column of Amberlite® IR 120 H⁺ form (Sigma-Aldrich Co.). The fractionation of humic substances was carried out as follows: aliquots of extracts were acidified to pH 2.0 with dilute H₂SO₄; the humic acids precipitated and were removed by centrifugation, while the fulvic acids corresponded to the supernatants (Bettany *et al.* 1980). The C content of humic and fulvic acids was determined by dichromate oxidation followed by titration with ferrous ammonium sulphate (Nelson and Sommers 1982).

Humification Ratio (HR%) representing the ratio HA+FA/TOC × 100, where HA and FA are purified humic and fulvic acids on PVP column (Petrucci *et al.* 1988) and TOC the total organic carbon, determined by the Springer and Klee method (Ministero delle Politiche Agricole Alimentari e Forestali 1994). The HR% parameter is proportional to the state of humification of the soil organic matter.

Phenols were extracted with distilled water, 1: 10 (w/v). Soil samples were shaken at 75 rev min⁻¹ for 20 h at room temperature and solutions were filtered through Whatman's no 1 paper. Total

water-soluble phenols (WSP) were measured by using the Folin-Ciocalteu reagent, following the Box method (1983). Tannic acid was used as a standard and the concentration of water-soluble phenols was expressed as tannic acid equivalents (mg TAE g⁻¹ d.w.) (Kuiters and Denneman 1987).

Experimental design

The soil (2 kg) was put in plastic pots and treated with OMW at five different concentrations (10, 25, 50, 75 and 100%). Untreated soil served as a control (0% OMW). All pots were supplemented with 120 mL of OMW at different concentrations or with distilled water (0% OMW). The pots with soil were then moistened up to 50% of water-holding capacity (WHC). Treatments were performed every week. All samples were incubated under controlled conditions of humidity and temperature, in a climatic chamber, using a regime of 16: 8 h light: dark, 24°C and a relative humidity of 70%. Each treatment was replicated five times. At 30 days' incubation, five pots of each treatment were taken and split into sub-samples for all analyses.

Microbial counts

Ten grams of each soil sample were added to 90 mL of 0.1% (w/v) sterile solution of sodium pyrophosphate and 1% glycerol, pH 7.0 (Elliot and Des Jardin 1999). After homogenization for 15 min at 1000 rpm, this solution was decimally diluted (10⁻¹ to 10⁻⁷) in a 0.85% NaCl sterile solution and aliquots of the resulting solutions plated on appropriate culture media. Tryptone Soya Broth (Oxoid Ltd, UK) diluted (3 g L⁻¹), containing cycloheximide (100 µg mL⁻¹, agent that inhibit fungal growth) and solidified with agar (15 g L⁻¹), adjusted to pH 7 and sterilized at 121°C for 20 min was used for total bacteria count. After incubation at 28°C for 28 days the colony forming units (CFU) were counted. Difco™ Malt Extract Agar containing chloramphenicol (inhibit bacteria growth) was used for fungal estimation (Picci and Nannipieri 2003). The incubation was performed at 28°C for 7 days. Actinomycetes were measured as CFU g⁻¹ of soil grown on starch casein medium (Wellington and Toth 1994), after the addition of nystatin (50 µg mL⁻¹, sterilized by filtration at 0.22 µm). The incubation was performed at 28°C for 14 days. Each soil sample was analysed in triplicate and the dilution series were plated in duplicate for each medium.

Statistical analysis

Data were subjected to analysis of variance (ANOVA) and treatment means were compared using the Tukey test at $p \leq 0.05$ (Sokal and Rohlf 1969).

RESULTS AND DISCUSSION

The results of selected chemical properties of the olive mill wastewaters stored in cement tanks during a period of 4 months (between October to February) suggest that some degradation and possible solid precipitation may occur during the stored period in open air. The degradation process was associated with the increase in pH (5.5), lower conductivity (4.9 dS m⁻¹), water soluble phenol and polyphenol content (5.47 and 11.33 mg TAE L⁻¹, respectively). The extremely low COD (10.5 g L⁻¹), and BOD (4.1 g L⁻¹) values seems not to be representative but rather reflects solid precipitation of some organic material. These results were confirmed also by the lower content of dry matter and ash in the stored OMW. Alteration of OMW properties are of practical relevance. In a real land application OMW should be stored for a longer period of at least several months before fully utilized. More changes may occur during a long storage period moreover if it is extended into the warm spring and early summer months, where the microbial processes are more intense (García *et al.* 1997; Muscolo and Sidari 2006). OMW storage could represent an efficient low cost pre-treatment before land application (Casa *et al.* 2003; Kistner *et al.* 2004; Lanciotti *et al.* 2005).

The sandy loam soil utilized in this trial had an alkaline

Table 2 Some chemical and physical properties of soil treated with different stored OMW concentrations.

	0%	10%	25%	50%	75%	100%
pH (H ₂ O)	8.10 a*	8.05 a	8.01 a	7.95 a	7.85 a	7.80 a
CO %	0.96 d	1.06 bc	1.04 c	1.11 b	1.26 a	1.31 a
SO %	1.65	1.82	1.78	1.91	2.16	2.25
Sand %	82 a					
Silt %	2 a	2 a	2 a	2 a	2 a	2 a
Clay %	16 a					
Texture	Sandy loam					
CaCO ₃ %	7 a	7 a	7 a	7 a	7 a	7 a
C.U. tot %	0.23 a	0.23 a	0.23 a	0.23 a	0.22 a	0.22 a
HR %	24 a	23 a	22 b	22 b	17 c	16 c
WSP mg TAE L ⁻¹	30 e	35 d	34 d	36 c	38 b	41 a

*values in the same row followed by the same letter are not significantly different ($p \leq 0.05$)

character (pH 8.1) (Table 2), and few content of organic carbon (0.96%). The results presented in this study demonstrate that the chemical properties of the soil amended with different concentrations of OMW were in part modified. After 30 days' incubation, OMW acidity (pH 5.5) has been neutralised by soil carbonate alkalinity (CaCO₃ 7%) and intrinsic buffering capacity of soil.

Soil organic carbon and water-soluble phenols increased (26 and 27%, respectively) after 100% OMW addition, no variations were recorded on humified organic carbon, probably for the short period of trial; consequently, a significant decrease in the humification ratio (HR%) was observed. These results are consistent with the results of Paredes *et al.* (1999) and Mekki *et al.* (2006). They have demonstrated that OMW increase soil organic matter, resulting in enhanced soil fertility.

The addition of OMW to the soil induced significant modifications in the soil microbial community. Bacteria decreased significantly up to 86% for 100% OMW treatment compared to control. This reduction is presumably due to the slight lowering soil pH caused by the addition of OMW (Fig. 1).

The growth of fungi was influenced by OMW treatment differently: no significant differences were found in the soil treated up to 50% OMW, a decrease in 50% was found in 100% OMW compared to control (Fig. 1). The fungal: bacteria ratio increased significantly. This suggests that the presence of OMW caused a shift between the proportion of fungi and bacteria, OMW represent an adequate substrate for fungi. The increase in fungi after application of OMW can help to degrade the phenolic and non phenolic aromatic pollutants in the OMW. The structure of the aromatic compounds in OMW presents obvious similarities to lignin components (Sanjust *et al.* 1991), and it is generally known that the ligninolytic enzymes are involved in the fungal degradation of polyphenols in OMW (Sayadi and Ellouz 1992). Fungal populations are known for their considerable diversity of depolymerising enzymes and for their resistance to recalcitrant substances (Dix and Webster, 1995). Evelyn *et al.* (2005) showed that fungi are the organisms principally responsible for lignin degradation in soil.

Under natural conditions, rarely are substances transformed by a single microbial species, rather a mixed flora is usually responsible for the conversions that occur. Because lignin is highly resistant, it protects cellulose against attack by most microbes, and it must be degraded by biological means before the cellulose can be utilized. A part some higher fungi such as the basidiomycetes, the actinomycetes are also capable of degrading substances with high resistance to microbial attack (Paul and Clark 1996), as some cellulose and lignin (Tuomela *et al.* 2000).

The addition of OMW to the soil caused a gradual decrease in actinomycetes CFU compared to control, until a maximum of 38% when the dose of OMW increased (Fig. 1).

The impact of OMW on soil microflora may be due to a temporary enrichment of the soil with nutritive solution, carbon source, and inhibiting components to some micro-

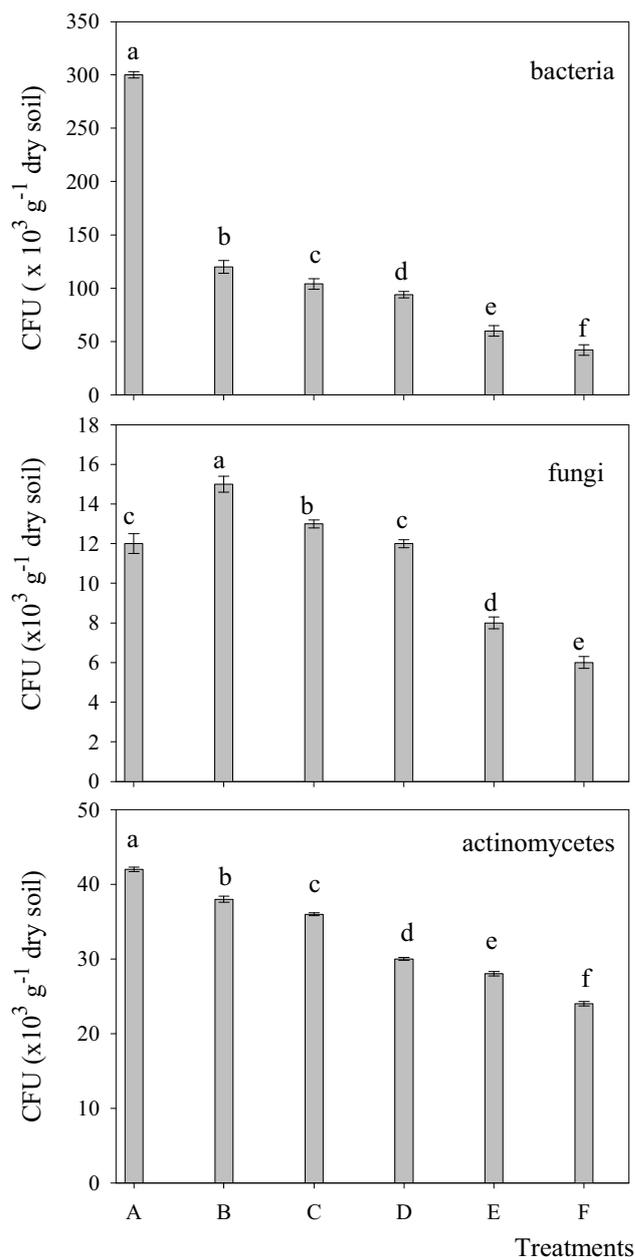


Fig. 1 Variation of CFU number of bacteria, fungi and actinomycetes in soil treated with different concentrations of OMW. A (0%), B (10%), C (25%), D (50%), E (75%), F (100%). Standard error of the mean is marked on the columns by bars. The same letters over bars indicate no significant difference according to Tukey's test ($p \leq 0.05$).

organisms. The results of this study show that OMW applied to soil up to 50% are not harmful to soil microflora in general. Yet, the possibility of inhibition of certain soil bac-

teria need to be further investigated. Possible changes in soil bacterial and fungal communities are not necessary detrimental to soil fertility. These changes may be the results of different fresh input of degradable carbon supplied with wastewater. Such changes are expected also with many types of soil organic additive as already reported by Stenström *et al.* (2001). Our results confirm previous findings (Alianello *et al.* 1998; Bååth and Anderson 2003; Mechri *et al.* 2007) evidencing that the microbial community responds to agronomic application of OMW with shift in key components of the biomass, and suggest that the changes in microbial structure may positively affect soil ecosystem leading to changes in the patterns of C, nutrient cycling and soil fertility.

CONCLUSIONS

Although the experiments presented here are limited by the laboratory controlled conditions, they may be suitable for assessing the short-term response of soil to an applied OMW. These results seem to confirm that the impact of OMW on soil properties and soil microbial communities depend on the relative amounts of beneficial and toxic compounds present, and soil had an intrinsic buffering capacity to resist to the applied perturbation.

The results of the present study indicate that the storage of OMW in open-air lagoons for a 4-month period and their opportune dilution with irrigation water before soil application may be a suitable inexpensive treatment to be adopted by small-sized olive mills for improving the quality of this material as soil amendment while minimizing the possible negative environmental impact.

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